

Detection of Mercury Chloride Acute Toxicity in Yellowfin Sea Bream (*Acanthopagrus latus*)

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Abstract: Toxicity tests allow the determination of pollution effects, providing direct evidence of the biological responses of marine organisms to contaminants. The 96-h LC₅₀ tests are conducted to measure the susceptibility and survival potential of organisms to particular toxic substances such as heavy metals. Hg²⁺ tested concentrations were 20, 50, 100, 200, 500, 1000, 2000, 5000 and 10000 µg/l, Groups of six male yellow fin sea bream (120 g) were exposed for 96 h to each of the Range Finding Test for LC50, in fiberglass tank equipped with aeration with 100 l of test medium. According to Range Finding Test (fifty percent of mortality between 500 and 1000) another tested concentration 550, 650, 750, 850 and 950 µg/l, Groups of six male yellow fin sea bream were exposed for 96 h to each of the LC₅₀ 96h for test solutions. 24, 48, 72 and 96 h LC₅₀ were 962.75, 886.48, 886.48 and 648.86 respectively. The 96 h NOEC, LOEC and LC₅₀ were 500, 550 and 648.86 µg/l respectively. LC₅₀ values indicated that mercury is more toxic to *A. latus*. LC50 obtained in the present study compare with corresponding values that have been published in the literature for other species of fish, show different LC₅₀ of mercury in different species and even different time, but what is important, is the lower value of LC₅₀ for *A. latus* as compare with most species and confirm sensitively of *A. latus* to low mercury doses.

Key words: LC50 • Pollution • Toxicity

INTRODUCTION

Aquatic ecosystems are typically monitored for pollution of heavy metals using biological assays, one of the best biological assays are Fishes. Fishes have been reported to accumulate heavy metals in their tissues several times above environment levels. Fishes have been used for many years to determine the pollution status of water and regarded as excellent biomarkers of pollution and especial metals in aquatic ecosystems.

Heavy metals have long been recognized as serious pollutants of the aquatic environment. They cause serious abnormality in metabolic, physiological and structural systems when present in high concentrations in the environment [1].

Mercury (Hg) is a liquid metal at ambient temperatures and pressures. Toxic effects of mercury and its compounds depend on the chemical form of it. Mercury forms salts in two ionic states mercury (I) and mercury (II). Mercury (II), or mercuric salts, is much more common in the environment than mercury (I) or mercurous

salts. These salts, if soluble in water, are bioavailable and considered toxic. Organic forms of mercury are generally more toxic to aquatic organisms than inorganic forms. HgCl₂ can be converted into highly toxic methyl mercury by methylation through chemical or biological processes. Mercury also forms organometallic compounds, many of which have oil and petrochemicals industry uses [2].

Mercury in fish was already recognized as a public health and ecological problem in the 1960's. It was commonly assumed that local point sources were the main sources and many studies focused on waters with nearby point source contamination, like creeks and estuaries [3], mercury chloride is the key form between the gaseous metal forms transported through atmosphere and the methylmercury form that bioaccumulates in organism. Once it enters into the organism, mercury can draw various immunotoxic effects.

Toxicity tests allow the determination of pollution effects, providing direct evidence of the biological responses of Fish and other aquatic animals to contaminants. Due to the fact that organisms from

different species vary in their sensitivity towards chemical substances, it is difficult to set standards for protection of species with regard to pollutants in the environment. Extrapolation from one species to another is, therefore, difficult if their relative sensitivities are not known [4].

The 96-h LC₅₀ tests are conducted to measure the susceptibility and survival potential of animals to particular toxic substances such as mercury. Higher LC₅₀ values are less toxic because greater concentrations are required to produce 50% mortality in organisms [5]. The heavy metals that are toxic to many organisms at very low concentrations and are never beneficial to living beings are mercury, cadmium and lead [6].

Yellowfin sea bream can easily be raised under laboratory conditions and it fulfills most of the requirements of a model species and is available throughout the year, so it was selected for our bioassay, so the present study was conducted to determine the acute toxicity of the heavy metal compound HgCl₂ in a statistic system to the marine fish *Acanthopagrus latus*.

MATERIAL AND METHOD

The experiments were conducted on Ninety six immature male yellow fine sea bream (average weight: 120 g) which were captured from Mahshahr creeks with hooks. Upon capture only healthy fish, as indicated by their activity and external appearance, were maintained alive on board in a fiberglass tank. After reaching shore, samples transferred to a 300-L aerated vat filled with sea water and maintained there until transporting to the laboratory. In laboratory, fish housed in a re-circulatory system (300-L tanks) equipped with physical/biological filters filled with seawater and aerated, from October to November. All experiments have done in Mariculture Research Station of the South Iranian Aquaculture Research Center, Mahshahr, Iran. All samples were acclimated for one weeks in a 15 aerated fiberglass tank containing 46 ppt saltwater maintained at 25 C under a constant 12:12 L:D photoperiod. Acclimatized fish were fed daily with a live feed (Fresh shrimp) and daily water quality and water parameters we checked. Dead fish were immediately removed with special plastic forceps to avoid possible deterioration of the water quality [7].

LC50 is the ambient aqueous chemical activity causes 50% mortality in an exposed population. These calculations are based on two important assumptions. The first assumption is that the exposure time associated with

the specified LC50 is sufficient to allow almost complete chemical equilibration between the fish and the water. The second assumption is that the specified LC50 is the minimum LC50 that kills the fish during the associated exposure interval. Fortunately, most reliable LC50 ' satisfy these two assumptions [6].

Hg²⁺ tested concentrations were 20, 50, 100, 200, 500, 1000, 2000, 5000 and 10000 µg/l, Groups of six male yellow fine sea bream (120 g) were exposed for 96 h to each of the Range Finding Test for LC50, in fiberglass tank equipped with aeration with 100 l of test medium. The control group was exposed to filtered sea water in similar conditions.

The bioassay was performed in a temperature of 25±1°C and under a natural photoperiod (12hL: 12hD) controlled room. Test medium was not renewed during the assay and no food was provided to the animals. Values of pH, temperature and salinity were measured at time 0, 24, 48, 72 and 96 h.

At the end of the bioassay [1], Range values were determined and according to that (Fifty percent of mortality between 500 and 1000) another tested concentration 550, 650, 750, 850 and 950 µg/l, groups of six male yellow fine sea bream (100 g) were exposed for 96 h to each of the LC50 96h for test solutions in same condition with Range Finding Test. At the end of the bioassay, LC50 96h values were determined [4].

LC value and standard error SE of LC were calculated following the probit procedure method as described by Wardlaw [8]. The LC_{10,30,50,70,90} values are derived using simple substitution probit of 10,30,50,70 and 90 respectively for probit of mortality in the regression equations of probit of mortality vs. mercury. The 95% confidence limits for LC₅₀ are estimated by using the formula LC₅₀ (95% CL) = LC50 ± 1.96 [SE (LC50)]. The SE of LC₅₀ is calculated from the formula:

$$SE(LC50) = \frac{1}{b\sqrt{pnw}}$$

Where: b=the slope of the mercury/probit response (regression) line; p=the number of mercury used, n = the number of animals in each group, w = the average weight of the observations [9] (Table 1).

Acute toxicity tests were carried out in order to calculate the 96h-LC50 for mercury in yellow fin sea bream, based on Wardlaw [8]. Mortality was recorded after 24,48,72 and 96h and LC50 values and its confidence limits(95%)were calculated by Boudou and Ribeyre [3]. The test was carried out in triplicate. Percentages of fish mortality were calculated for each mercury concentration at 24, 48, 72 and 96 h of exposure.

Table 1: The 95% confidence limits for LC₅₀ of yellowfin sea bream

Concentration (µg/l)	24h	48h	72h	96h
b	0.012	0.002	0.002	0.009
p	5	5	5	5
n	6	6	6	6
W	120	120	120	120
SE	1.38	8.33	8.33	1.85
95% CL	3.5868	16.3268	16.3268	3.626

RESULTS

There was 100% mortality at 10000 µg/l concentration within the first 4h after dosing and 100% mortality at 5000 µg/l within the 14h whereas 100% mortality for 2000 µg/l was 42h and for 1000 µg/l was 54h.

The mortality of yellowfin sea bream for mercury chloride doses 20, 50, 100, 200, 500, 1000, 2000, 5000 and 10000 µg/l were examined during the exposure times at 24, 48, 72 and 96 h for Range Finding Test (Table 2). Fish exposed during the period 24-96h had significantly (P<0.05) increased number of dead yellowfin sea bream with increasing concentration. There were significant (P<0.05) differences in number of dead fish between the duration 24-96 in each.

After finding this fact that main range is between 500-1000 (Because of no mortality at 500 µg/l and 100% mortality at 1000 µg/l), the mortality of yellowfin sea bream for mercury chloride doses 550, 650, 750, 850 and 950 µg/l were examined during the exposure times at 24, 48, 72 and 96 h for LC50 Test (Table 3).

Median lethal concentrations of 10, 30, 50, 70 and 90% test are in Table 4. Physicochemical parameters of test water are in Table 5.

Mortality percentage of Range Finding Test and LC50 experiment are in Figure 1, respectively, however sigmoid probit analyses and regression lines of probit are in Figures 2, 3 and 4, respectively.

Table 2: Cumulative mortality of yellowfin sea bream (n=6, each concentration) at Range Finding Test

Concentration (µg/l)	No. of dead yellowfin sea bream			
	24h	48h	72h	96h
Control	-	-	-	-
20	-	-	-	-
50	-	-	-	-
100	-	-	-	-
200	-	-	-	-
500	-	-	-	-
1000	1	3	6	6
2000	2	6	6	6
5000	6	6	6	6
10000	6	6	6	6

Table 3: Cumulative mortality of yellowfin sea bream (n=6, each concentration) at LC50 test

Concentration (µg/l)	No. of dead yellowfin sea bream			
	24h	48h	72h	96h
Control	-	-	-	-
550	-	-	1	1
650	-	1	2	3
750	1	3	5	5
850	1	3	6	6
950	1	3	6	6

Table 4: Lethal concentrations (LC_{1.99}) of mercuric chloride depending on time (24-96h) for *A. latus*

Point	Concentration (µg/l)		(95 % of confidence limits)	
	24h	48h	72h	96h
LC ₃₀	919.4132	705.6551	705.6551	594.8041
LC ₄₀	941.8181	799.1379	799.1379	622.7525
LC ₅₀	962.7520	886.4827	886.4827	648.8659
LC ₆₀	983.6859	973.8275	973.8275	674.9793
LC ₇₀	1026.0909	1007.3103	1007.3103	702.9278

Table 5: Physicochemical parameters of test water

Parameters	
Temperature (°C)	25 ± 1
pH	7.8 ± 0.1
Salinity	46±1

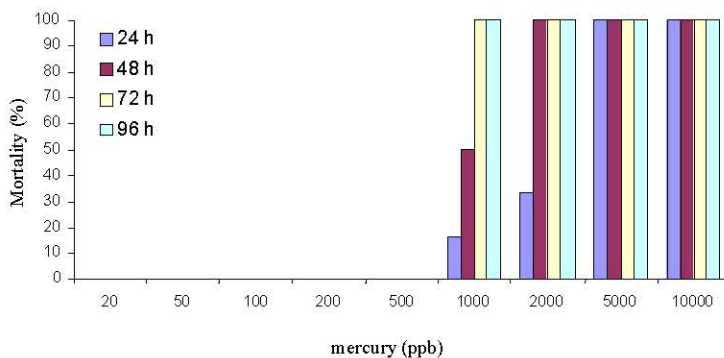


Fig. 1: The column mercury-response (mortality) for *A. Latus* in the Range Finding Test experiment

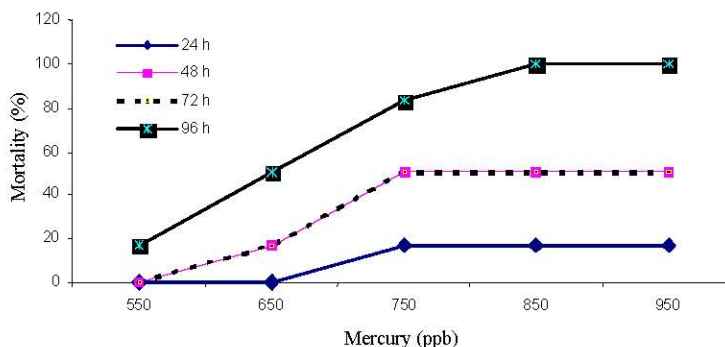


Fig. 2: The sigmoid mercury-response (mortality) curve for *A. Latus* in the LC₅₀ experiment

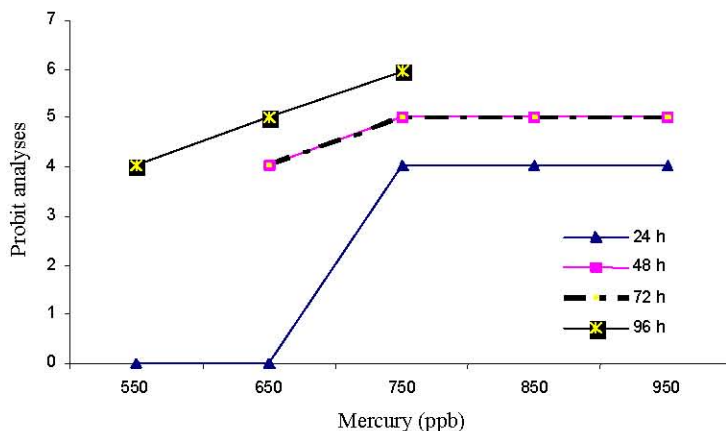


Fig. 3: The sigmoid probit analyses curve for *A. Latus* in the LC₅₀ experiment

Toxicity Testing Statistical Endpoints are in tow part:
 1-Hypothesis Testing: is there a statistically significant difference between the mean response in the treatments and mean response in control or reference sample?
 LOEC: Lowest Observed Effect Concentration; NOEC:

No Observed Effect Concentration. 2-Point Estimates: what toxicant concentration will cause a specific effect on the test population? LC50: the median Lethal Concentration. Our result for Toxicity Testing Statistical Endpoints is in Fig 5.

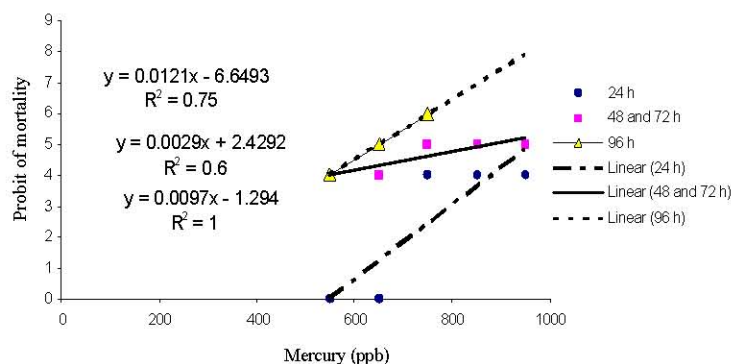


Fig. 4: Probit of mortality versus mercury regression lines for *A. latus* in the LC₅₀ experiment. Also depicted are the regression equations and R^2 values. Probit values used are derived from Fig. 3.

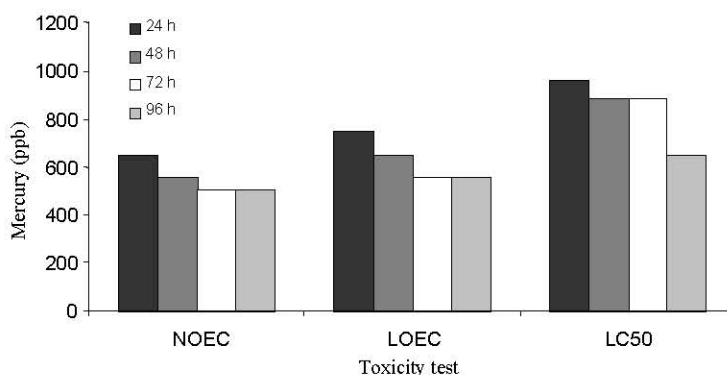


Fig. 5: Toxicity testing statistical endpoints in yellowfin sea bream

DISCUSSION

Mercury levels can be alternated with different conditions. Factors influencing mercury levels can be divided into exogenous (Characteristics of the water body) and endogenous (Characteristic of the individuals or species). Exogenous factors include pH, temperature and salinity. Endogenous factors include species, habitat and food preferences, metabolic rate, age, growth rate, size, mass and diet, so all Fish were same in sexually, size, weight, natural location, moreover all laboratorial parameters were fix for all Fish.

Mercury is one of the concern metals in aquaculture and has 10-40 $\mu\text{g/l}$ of LC₅₀ with only 1 $\mu\text{g/l}$ for safe levels, whereas LC₅₀ value for other heavy metals is higher than mercury (Cadmium 80-420, cooper 20-100, zinc 1000-10000, lead 1000-40000 $\mu\text{g/l}$). also the 96h LC₅₀ for the juvenile trout as 11 $\mu\text{g/l}$ (95% CI = 9.2-11.9 $\mu\text{g/l}$) [7].

The 96-h LC₅₀ value for catfish exposed to Hg^{2+} under static test was determined to be 570 $\mu\text{g/l}$ [10]. The 96-h LC₅₀ value of mercury chloride for chub was found as 205 $\mu\text{g/l}$ and 96-h LC₅₀ for trout 814 $\mu\text{g/l}$ [11]. On the estuarine fish *Pomatoschistus microps*, LC₅₀ of copper

and mercury at 96 h were 568 $\mu\text{g/l}$ and 62 $\mu\text{g/l}$, respectively [12]. The concentrations of trace metals that resulted in mortality of *H. rubra* were investigated by exposing juveniles to acute concentrations of Cu, Zn, Hg and Cd for 96hr. Hg resulted in more sudden mortality rate after 24hr exposure compared to Cu yet produced a 96hr LC₅₀ of 173 $\mu\text{g Hg/L}$ [13], this result show except trout, all smaller fishes have low level of LC₅₀ in compare to yellowfin sea bream is higher than all fishes.

Many aquatic species show vast range of LC₅₀ for mercury chloride, which for saltwater fish was 36 $\mu\text{g/l}$ (juvenile spot) to 1678 $\mu\text{g/l}$ (flounder), that was higher than saltwater invertebrate 3.5 $\mu\text{g/l}$ (mysid shrimp) to 400 $\mu\text{g/l}$ (soft clam) (12-13). This result emphases that yellowfin sea bream is sensitive to mercury chloride and have low LC₅₀ value.

Other study show different results, for example FAO/UNEP [14] find that, the 96-h LC₅₀ values of mercury chloride are for cat fish 350 $\mu\text{g/l}$, rainbow trout 220 $\mu\text{g/l}$, striped bass 90 $\mu\text{g/l}$ and brook trout 75 $\mu\text{g/l}$.

The 96-h LC₅₀ values of mercury chloride 37 $\mu\text{g/l}$ for fathead minnow, 160 $\mu\text{g/l}$ for bluegill sunfish, 903 $\mu\text{g/l}$ for rainbow trout, 200 $\mu\text{g/l}$ for rainbow trout and lower in

invertebrate, 2 µg/l for crayfish, 5 µg/l for cladocera, 10 µg/l for Gammarus, 5 µg/l for blue mussel, 15 µg/l for prawn and 3 µg/l for limpet (Eisler, 1987). For mercury, 96 h LC50 values of 75 µg/l for the catfish (*Sarothodon mossambicus*), 33 µg/l for the rainbow trout (*Salmo gairdneri*), 110 µg/l for the banded killifish (*Fundulus diaphanous*) and 90 µg/l for the striped bass (*Roccus saxatilis*) were found (15-17).

The susceptibility of fish to a particular heavy metal is a very important factor for LC50 values. The fish that is highly susceptible to the toxicity of one metal may be less or non-susceptible to the toxicity of another metal at the same concentration of that metal in the milieu. Similarly, the metal which is highly toxic to one organism at low concentration may be less or non-toxic to other organism at the same or even higher concentration, so the LC₅₀ values reported in the present study for HgCl₂ were lower than the values reported by Agarwal [15] for the *Channa punctatus* (Bloch) at 48, 72 and 96 h. He reported LC₅₀ values of 2.512, 2.291 and 2.113 mg/L, respectively, at 48, 72 and 96 h. however, the present values, are higher than those of Khangarot and Ray [16]: 0.432 and 0.314 mg/L, respectively, at 72 and 96h in *Channa marulius*.

The acute toxicity of HgCl₂ increases with increase in temperature [17] and similar trends for other metals [18]. Also other researcher observed that the toxicity of copper abruptly decreased with an increase in pH of the Cu-containing medium [16]. Acute toxicity studies are the very first step in determining the water quality requirements of fish. These studies obviously reveal the toxicant concentrations (LC₅₀) that cause fish mortality even at short exposure. Therefore, studies demonstrating the sensitivity of genotoxic effects of heavy metals in aquatic organisms, particularly in fish are needed. Thus, it can be concluded from the present study that fish are highly sensitive to HgCl₂ and their mortality rate is dose dependent.

Comparison of values reported earlier with those obtained in the present study may not be meaningful because various factors may influence bioassay techniques like differences in fish (e.g., species, weight, size) and other environmental factors (Temperature, variations in pH of the water, total hardness of water, dissolved oxygen). variability in acute toxicity even in a single species and single toxicant depending on the size, age and condition of the test species along with experimental factors [19]. The differences in acute toxicity may be due to changes in water quality and test species [20].

Chronic toxicity values are much lower than acute values and highlight the adverse effects of relatively low concentrations of mercury in water (i.e., <1 µg/L). In aquatic toxicology, if LC50 concentration is smaller than 1000 µg/l, the chemical is highly toxic and if between 1000-10000 µg/l, then it is considered to be moderately toxic [21], therefore it is reported here that mercury chloride is highly toxic to yellowfin sea bream and may cause many damage in this fish.

The fish exposed to metal can compensate for the stressors. If it cannot successfully compensate for stressor effects, an altered physiological stage may be reached in which the organism continues to function and, in extreme cases, the acclimation response may be exhausted with a subsequent effect on fitness [22]. In the present study, LC50 values indicated that mercury is more toxic to *A. latus*. LC50 obtained in the present study (650 µg/l) compare with corresponding values that have been published in the literature for other species of fish, show different LC50 of mercury in different species and even different time, but what is important, lower value of LC50 for *A. latus* compare with most species and confirm sensitively of *A. latus* to low mercury doses.

It was concluded that LC₅₀ of mercury chloride for *A. latus* is 650 µg/l, this species is highly sensitive to mercury and probably to other environmental pollution, therefore progress of human induced pollutant may cause serious damage to population of this important fish.

ACKNOWLEDGEMENTS

The authors are thankful to the Director and Staff at the Mariculture Research Station, Mahshahr, Iran for providing necessary facilities for the experiment and the University of Marine Science and Technology, Khoramshahr, Iran, for support during the tenure of this project. We would like to express our special thanks to Dr. Movahedinia for assistance during the trials.

REFERENCES

1. Tort, L., P. Torres and R. Flos, 1987. Effects on dogfish haematology and liver composition after acute copper exposure. *Comp. Biochem. Physiol.*, 87: 349-353.
2. Boening, D., 2000. Ecological effects, transport and fate of mercury: A general review. *Chemosphere*, 40: 13-51.
3. Boudou, A. and F. Ribeyre, 1997. Aquatic ecotoxicology from the ecosystem to the cellular and molecular levels. *Environ. Health Perspect*, 105: 21-35.

4. Van Straalen, N.M., P. Leeuwangh and P.M.B. Stortelder, 1994. Ecotoxicology of soil organisms. CRC Press.
5. Eaton, A.D., L.S. Clesceri, A.E. Greenberg and M.A.H. Franson, 1995. Standard Methods for the Examination of Water and Waste Water. American Public Health Association Press.
6. Hilmy, A.M., M.B. Shabana and A.Y. Dabees, 1985. Bioaccumulation of cadmium: Toxicity in *Mugil cephalus*. Comparative Biochem. Physiol., 81: 139-143.
7. Gooley, G.J., F.M. Gavine and L. Olsen, 2006. Biological Systems to Improve Quality and Productivity of Recycled Urban Wastewater. A Joint Project of: Department of Primary Industries, Victoria.
8. Wardlaw, A.C., 1985. Practical Statistics for Experimental Biologists. Wiley Press, pp: 104-110.
9. Hotos, G.N. and N. Vlahos, 1998. Salinity tolerance of *Mugil cephalus* and *Chelon labrosus* Pisces: Mugilidae/fry in experimental conditions. Aquaculture, 167: 329-338.
10. Elia, A.C., L. Mantilacci, M.I. Taticchi and R. Galarini, 2000. Effects of mercury on glutathione and glutathione-dependent enzymes in catfish (*Ictalurus melas* R.). In: Markert, B., Friese, K. (Eds.), Trace Elements - Their Distribution and Effects in the Environment: Trace Metals in the Environment. Elsevier Science Press, Amsterdam, pp: 411-421.
11. Verep, B., E. Sibel Besli, I. Altionk and C. Mutlu, 2007. Assessment of Mercuric chloride toxicity on Rainbow trouts and cubs. Pakistan J. Biol. Sci., 10: 1098-1102.
12. Vieira, L.R., C. Gravato, A. Soares, F. Morgado and L. Guilhermino, 2009. Acute effects of copper and mercury on the estuarine fish *Pomatoschistus microps*: Linking biomarkers to behavior. Chemosphere, 76: 1416-1427.
13. Gorski, K., 2007. The effects of trace metals on the Australian abalone, *Haliotis rubra* Jacqueline. PhD thesis. RMIT University, pp: 84.
14. FAO/UNEP, 1996. Operation of the prior informed consent procedure for banned or severely restricted chemicals in international trade. Joint FAO/UNEP program, Rome, Geneva, Amended.
15. Agarwal, S.K., 1991. Bioassay evaluation of acute toxicity levels of mercuric chloride to an air-breathing fish *Channa punctatus* (Bloch): mortality and behavior study. J. Environ. Biol., 12: 99-106.
16. Khangarot, B.S. and P.K. Ray, 1987. Response of a freshwater Ostracod (*Cypris subglobosa* Sowerby) exposed to copper at different pH levels. Acta Hydrochemistry Hydrobiol., 15: 553-558.
17. Rathore, R.S. and B.S. Khangarot, 2002. Effect of temperature on the sensitivity of sludge worm *Tubifex tubifex* (Muller) to selected heavy metals. Ecotoxicol. Environ. Safely, 53: 27-36.
18. Cairns, J., J. Buikema, A.L. Heath and A.G. Parker, 1981. Effects of temperature on aquatic organism sensitive to selected chemicals. Water Resources Resource Center Bulletin, 106: 1-88.
19. Sprague, J.B., 1969. Measurement of pollutant toxicity to fish: I. Bioassay methods for acute toxicity. Water Resources, 3: 793-821.
20. Gupta, P.K., B.S. Khangarot and V.S. Durve, 1981. The temperature dependence of the acute toxicity of copper to a freshwater pond snail. *Viviparus bengalensis* L. Hydrobiologia, 83: 461-464.
21. Louis, A.H., L.W. Diana, H. Patricia and R.S. Elizabeth, 1996. Pesticides and Aquatic Animals, [C]. Virginia Cooperative Extension, Virginia State University, Virginia, pp: 24.
22. Mayer, F.L., D.J. Versteeg, M.J. McKee and L.C. Folmar, 1992. Physiological and non-specific biomarkers. In: Biomarkers: Biochemical, physiological and histological markers of anthropogenic stress. Huggett, R.J., Kimerle, R.A., Mehrle, P.M. Jr., Bergman, H.L. Lewis Publishers, Boca Raton, pp: 5-85.